



Valorisation of bauxite residue as filter material to upgrade phosphorus removal in small wastewater treatment plants

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Context: phosphorus pollution

Phosphorus (P) is an **essential nutrient** for biomass growth in aquatic ecosystems.

Excessive intake of P in water bodies may lead:

- **Abnormal growth of algae and aquatic plants (algal bloom);**
- **Degradation of water quality (eutrophication).**



Maine-et-Loire (France)



Lake Winnipeg (Canada)

Images from Internet

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Treatment requirements for small and medium (10000-100000 P.E.) wastewater treatment plants (WWTP) in sensitive areas (directives 91/271/EEC and 2000/60/EC):

- Total phosphorus concentration (TP): **2 mg P/L;**
- Minimum percentage of reduction: **80%;**
- **National and local requirements are often stricter (even < 0.5 mg P/L)!**

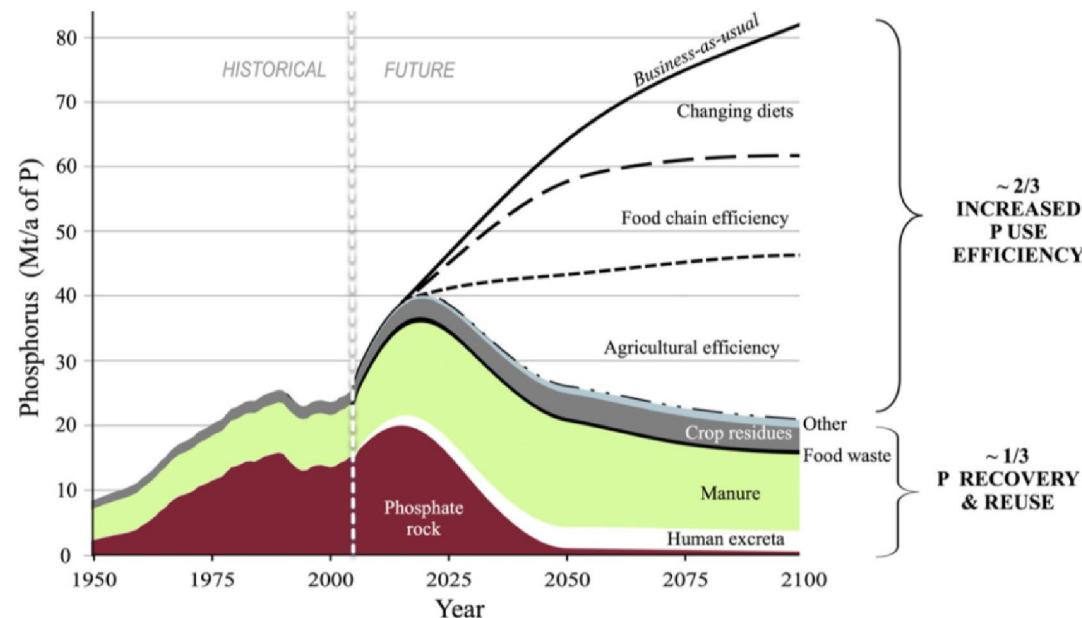
Domestic wastewater in Western Europe (Comber *et al.*, 2013; Boutin et Eme, 2017):

- **Per capita loadings of P: 2.0-2.6 g P per capita per day;**
- **TP concentration (fresh wastewater): 10-18 mg P/L ;**
- **The need to develop low cost techniques to treat P, especially for small WWTPs.**

Context: shortage of natural resources

Nowadays fertilizer production industry strongly depends on natural deposits of P such as phosphate rocks. Prospective studies indicate that (Cordell *et al.*, 2011):

- The peak of P production from phosphate rocks will occur around 2020;
- P is very likely to become a critical resource by 2050;
- There is an urgent need to identify alternative renewable P resources.



Scenario of long term phosphorus demand (Cordell *et al.*, 2011)

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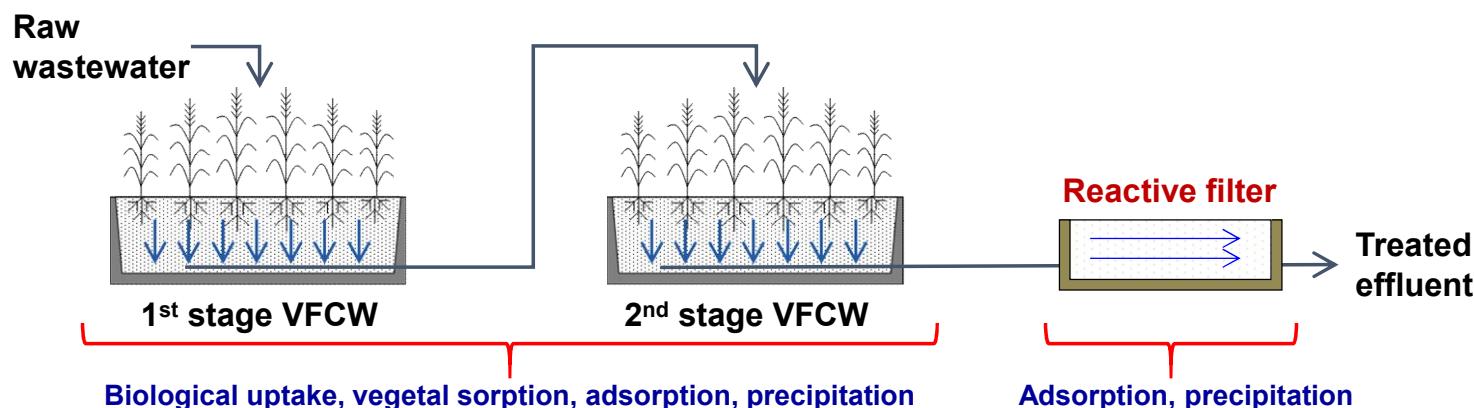
P retention and recovery from domestic wastewater represents a promising strategy to (Tarayre *et al.*, 2016; Cieślik and Konieczka, 2017):

- Reduce P supply to sensitive ecosystems (risk of eutrophication);
- Overcome the shortage of natural deposits of P (e.g. phosphate rocks);
- Maximum potential of P recovery: 0.7-0.9 kg P per capita per year.

Context: P treatment in small WWTP

Most common treatment systems for small communities in France (< 2000 P.E.):

- Two stage vertical flow reed planted constructed wetland (VFCW):



Two stage VFCWs in France provide (Paing et al., 2015):

- High removal (> 90%): COD, suspended solid (TSS), and Kjeldahl nitrogen (TKN);
- Poor removal (< 30%): nitrate (N-NO₃) and total phosphorus (TP);
- Addition of separate filter units containing materials with high affinity for P binding.

Context: reactive materials

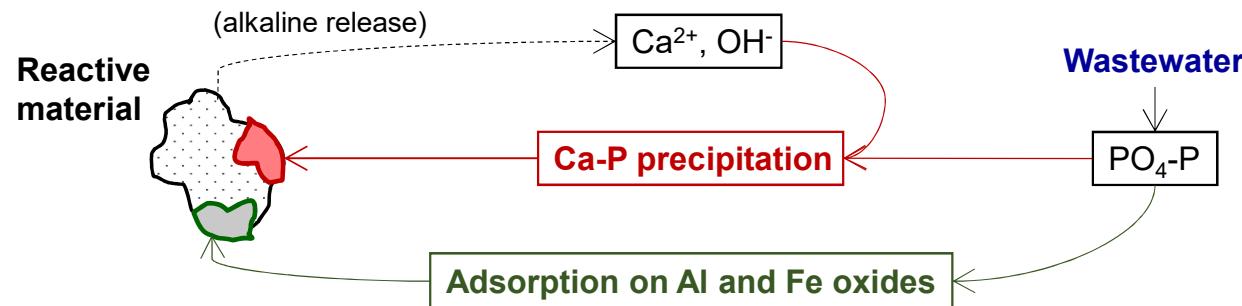
Reactive materials tested for P binding (Vohla *et al.*, 2011; Bacelo *et al.*, 2020):

- Natural materials: **limestone, zeolite, iron rich sand, etc.;**
- Man made: **Filtralite®, Phosphorite®, Polonite®, etc.;**
- Industrial byproducts and waste: **steel slag, fly ash, bauxite residue, etc.**

Most of these materials present **high Ca, Al and/or Fe content.**

Main mechanisms of P binding (Barca *et al.*, 2012):

- **Precipitation of Ca-P complexes followed by crystallization on mineral surface;**
- **Adsorption on Al and Fe oxides and hydroxides.**



Carbonated bauxite residue as filter material

Project BAUXFILTER (ALTEO, LabEx DRIIHM OHM-BMP, 2018-2019):

- Laboratory M2P2, group Waste and Wastewater Treatment, Aix-en-Provence;
- INERIS-ARDEVIE, Aix-en-Provence;
- Company ALTEO, Gardanne (Provence, France).

Bauxite residue: waste of aluminum industry (also known as red mud):

- Worldwide production (Prajapati et al., 2016): **90 million tons per year**;
- Chemical composition (ALTEO): **Fe₂O₃ (50%), Al₂O₃ (14%), CaO (5,5%), Na₂O (3,5%)**;
- High content of NaOH: **high pH leachates**;
- Carbonated bauxite residue (CBR): **treated by addition of gypsum to reduce pH < 8.5.**



Bauxite residue storage area of Gardanne with Sainte-Victoire Mountain (Provence, France)



Vegetation growth on CBR, field-study (1000 m²) (Hennebert et al., 2017)

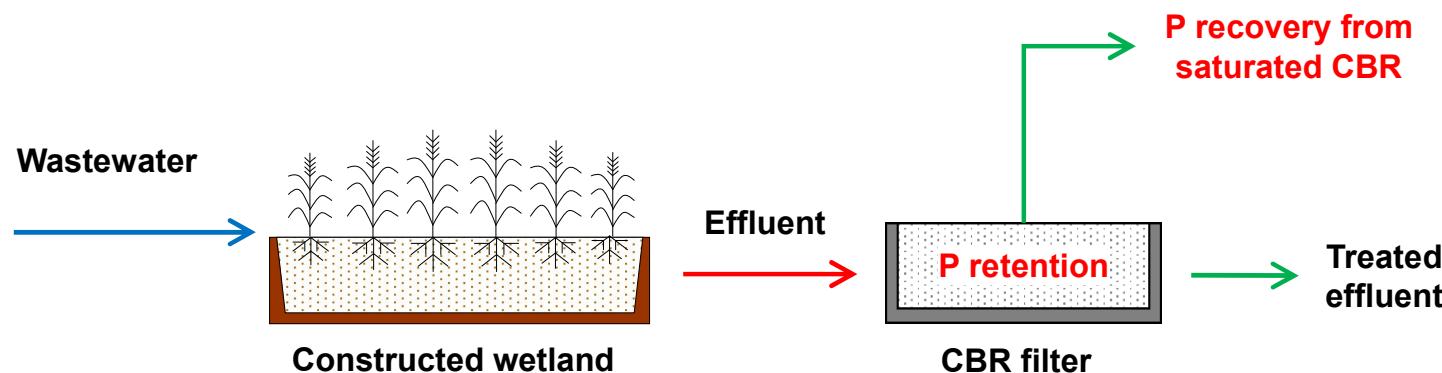
The project BAUXFILTER

Aim of the project:

- Developing the use of CBR filters to remove P from the effluents of small WWTPs.

Main challenges:

- To reduce P supply to receiving waters;
- To valorize an industrial waste as filter material;
- To retain and recover P from wastewater.

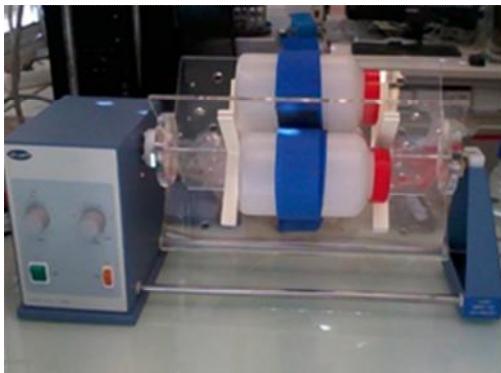


Main objectives and approach

Multi-scale approach involving experiments at different scales of investigation:

- I. Batch sorption experiments: **kinetics and equilibrium capacities of P sorption;**
 - II. Lab-scale column experiments: **P removal performance under dynamic conditions;**
 - III. Chemical extractions: **main mechanisms of P removal achieved by CBR.**
- Integration of results and design of a pilot scale filter.

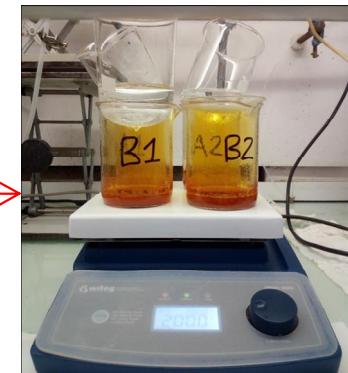
Batch sorption experiments



Column experiments



Chemical extractions

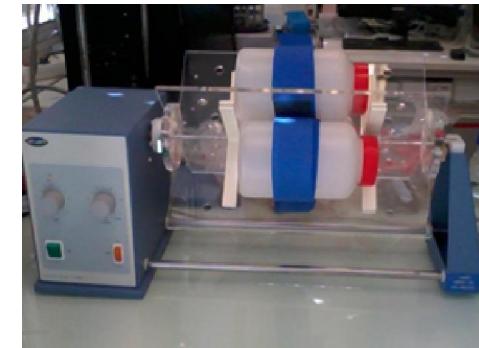


Integration of results

Material & methods: batch experiments

Batch kinetic experiments: to determine the effect of different wastewater composition on equilibrium capacities and rate constants of P sorption:

- Ratio liquid to solid (ASTM D 4646): **20 L/kg**;
- Initial volume of solutions: **0.7 L**;
- Agitation mode: **rotary agitation at 2.5 rpm**;
- Room temperature: **$20 \pm 2^\circ\text{C}$** ;
- Water samples taken at: **0.5, 1, 2, 4, 6, and 24 h**.

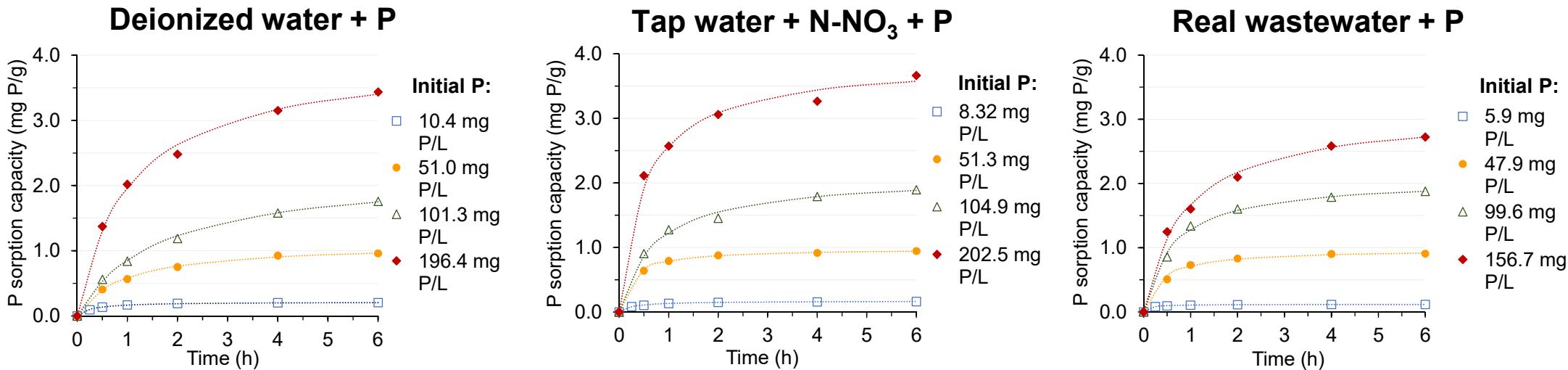


Solutions: 3 different water matrix at 4 different initial P:

- Deionized water plus P: **10, 50, 100, and 200 mg P/L**;
- Tap water plus 40 mg N-NO₃/L plus P: **10, 50, 100, and 200 mg P/L**;
- Real wastewater* plus P: **10, 50, 100, and 150 mg P/L**.

*Effluent from the two stage VFCW of Rougiers (Var, France), 1500 P.E..

Results & discussion: batch experiments



Pseudo 2nd order model (Ho and McKay, 1998):

$$\frac{dq_t}{dt} = k_2 \cdot (q_e - q_t)^2 \quad \rightarrow \quad \frac{t}{q_t} = \frac{1}{k_2 q_e^2} + \frac{1}{q_e} t$$

- q_e : equilibrium sorption capacity (mg P/g CBR);
- q_t : sorption capacity at time t (mg P/g CBR);
- k_2 : rate constant of pseudo-second order ($\text{g mg}^{-1} \text{ h}^{-1}$).

One or more reactants become limiting: process controlled by the reaction.

Results & discussion: batch experiments

Deionized water + P

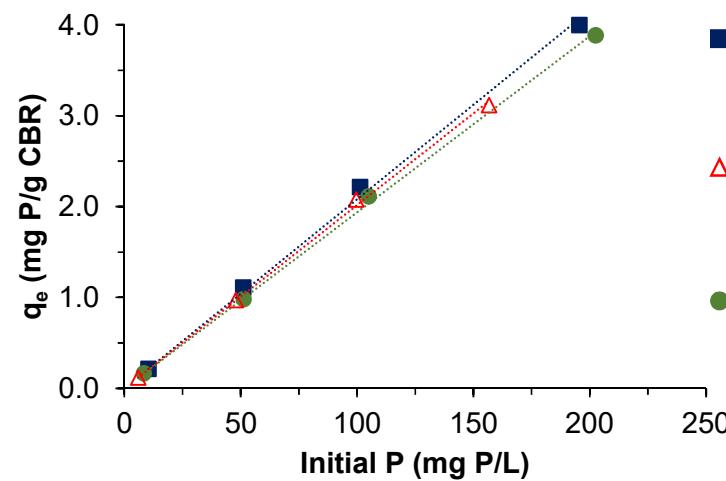
Initial P (mg P/L)	K_2 (g/(mg*h))	q_e (mg/g)	R^2 (-)
10.4	16.09	0.22	0.999
51.0	0.99	1.16	0.999
101.3	0.28	2.22	0.998
196.4	0.24	3.99	0.997

Tap water + N-NO₃ + P

Initial P (mg P/L)	K_2 (g/(mg*h))	q_e (mg/g)	R^2 (-)
8.32	21.48	0.17	0.999
51.3	4.01	0.98	0.999
104.9	0.65	2.12	0.997
202.5	0.50	3.88	0.995

Real wastewater + P

Initial P (mg P/L)	K_2 (g/(mg*h))	q_e (mg/g)	R^2 (-)
5.9	75.08	0.12	0.999
47.9	2.78	0.97	0.999
99.6	0.77	2.08	0.999
156.7	0.36	3.12	0.998



■ Deionized water + P:

$$q_e = 0.0208 \cdot \text{Initial P} \quad (R^2 = 0.998)$$

△ Tap water + N-NO₃ + P:

$$q_e = 0.0194 \cdot \text{Initial P} \quad (R^2 = 0.999)$$

● Real wastewater + P:

$$q_e = 0.0202 \cdot \text{Initial P} \quad (R^2 = 0.999)$$

- q_e increases according to the increase in initial P: **saturation capacity not achieved;**
- Different water matrix did not seem to significantly affect kinetics and capacity of P sorption.

Material & methods: column experiments

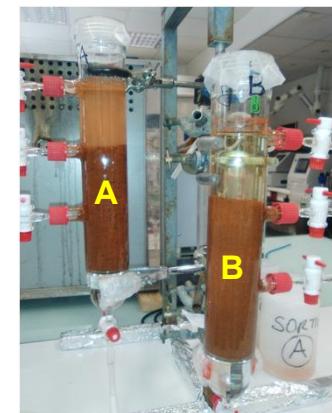
Main objectives: to determine and describe the effect of **aerobic and anoxic conditions** on:

- P removal performances;
- P removal mechanisms.

Two CBR columns were continuously fed according to a HRT_v of 1 day for the full period of 5 months of operation:

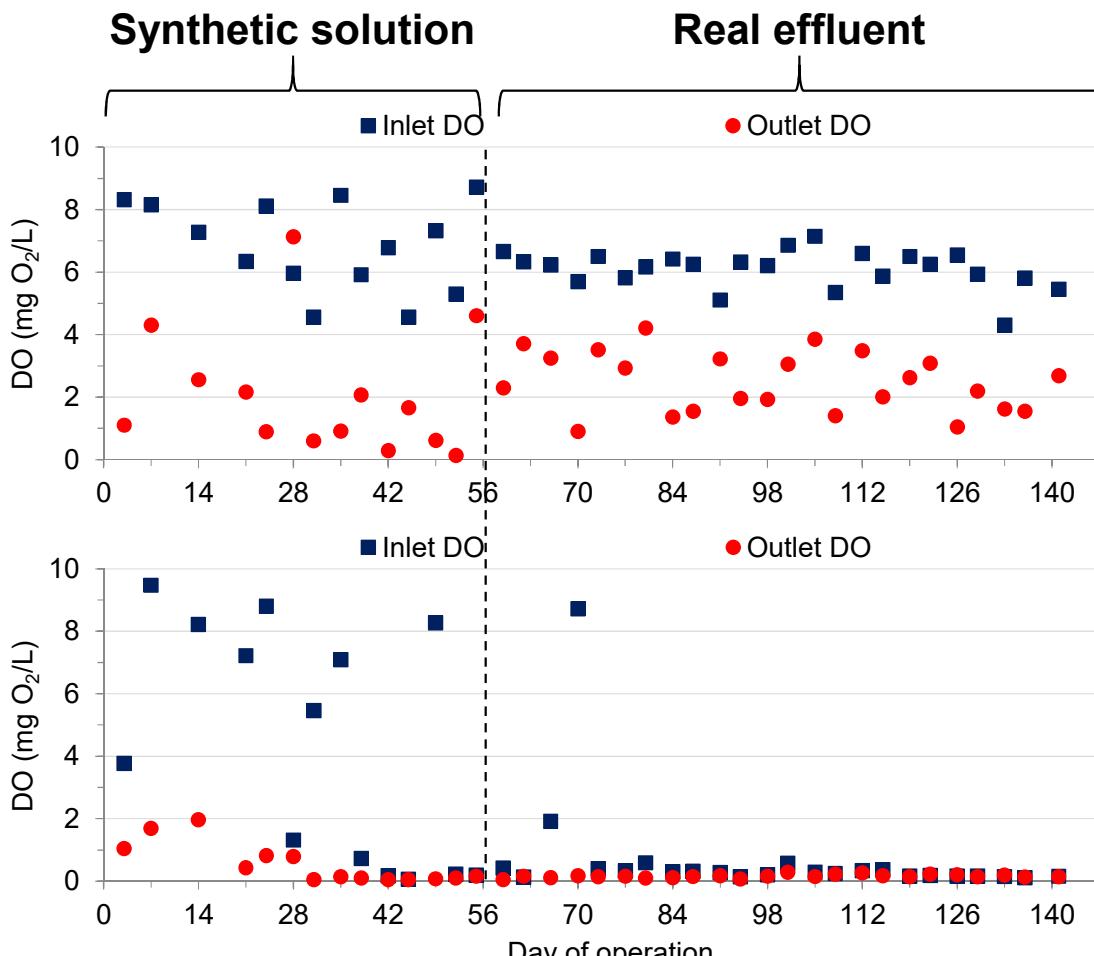
- Day 1 to 54: **synthetic solution**:
 - Column A: tap water + 10 mg P/L + 40 mg N/L (KNO_3);
 - Column B: tap water + 10 mg P/L + 40 mg N/L (KNO_3) + 500 mg COD/L (glucose);
- Day 55 to 140: **real effluent from a small WWTP***:
 - Column A: **raw real effluent**;
 - Column B: **real effluent + 500 mg COD/L**.

Glass columns (0.5 L)



*Effluent from the two stage VFCW of Rougiers (Var, France), 1500 P.E..

Results & discussion: DO concentrations



Column A: aerobic conditions

Synthetic solution (day 1-54):

- Outlet DO: $1.9 \pm 1.9 \text{ mg O}_2/\text{L}$

Real effluent (day 55-140):

- Outlet DO: $2.6 \pm 1.0 \text{ mg O}_2/\text{L}$

Column B: anoxic conditions

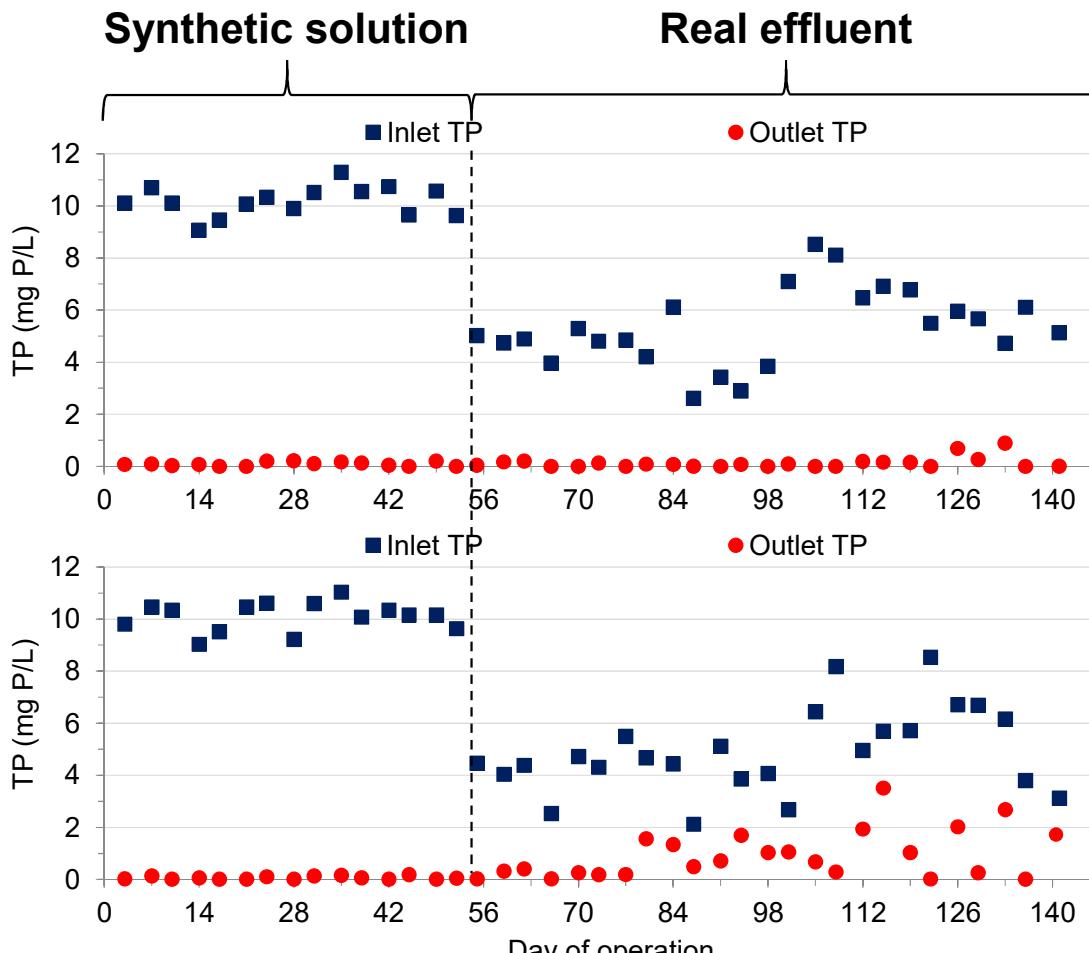
Synthetic solution (day 1-54):

- Outlet DO: $0.6 \pm 0.6 \text{ mg O}_2/\text{L}$

Real effluent (day 55-140):

- Outlet DO: $0.2 \pm 0.1 \text{ mg O}_2/\text{L}$

Results & discussion: TP removal



Column A: aerobic

Synthetic solution (day 1-54):

- Inlet TP: 10.3 ± 0.5 mg P/L
- Outlet TP: 0.1 ± 0.1 mg P/L

Real effluent (day 55-140):

- Inlet TP: 5.5 ± 1.5 mg P/L
- Outlet TP: 0.1 ± 0.1 mg P/L

Column B: anoxic

Synthetic solution (day 1-54):

- Inlet TP: 10.3 ± 0.5 mg P/L
- Outlet TP: 0.1 ± 0.1 mg P/L

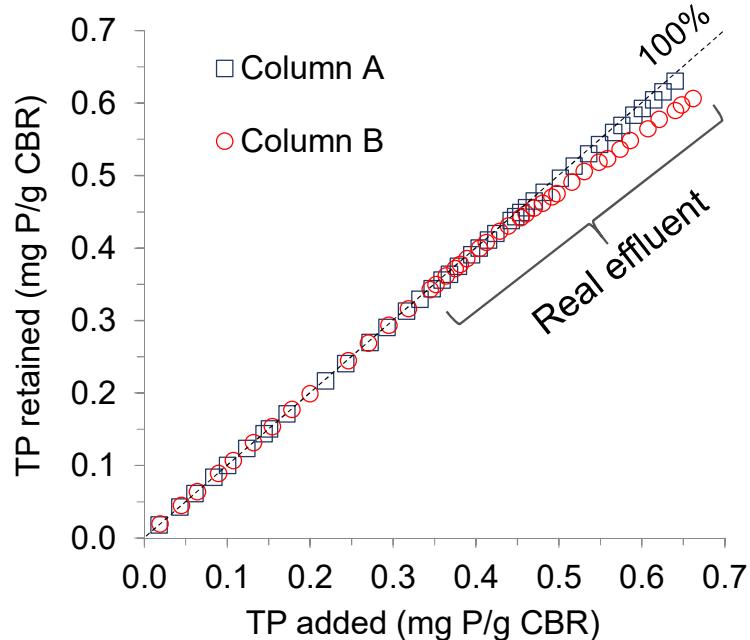
Real effluent (day 55-140):

- Inlet TP: 5.5 ± 1.7 mg P/L
- Outlet TP: 0.9 ± 0.9 mg P/L

Journées Scientifiques :
"Bauxite Résidues"

Aix-en-Provence, du 08 au 09 juin 2021

Results & discussion: TP retention capacity

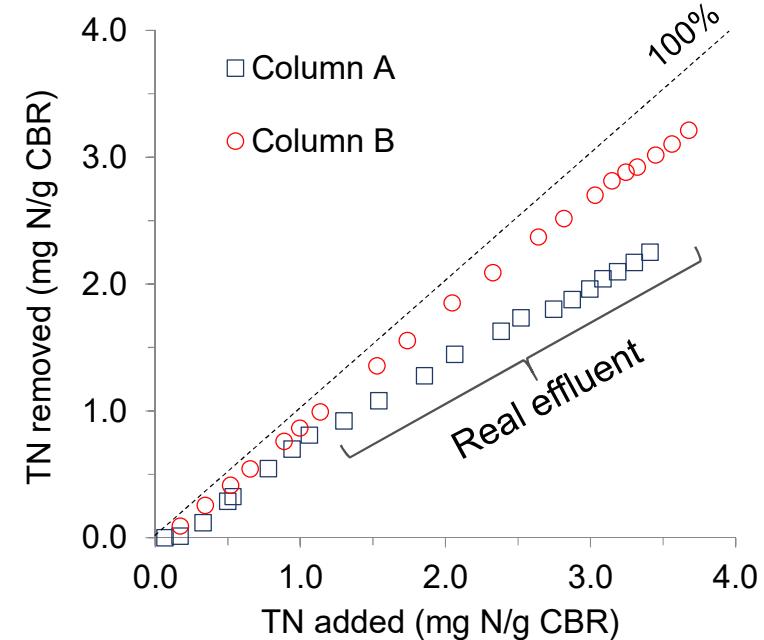


TP retention capacities over 140 days:

- Column A: **0.63 mg P/g CBR**
- Column B: **0.61 mg P/g CBR**

TP retention efficiency over 140 days:

- Column A: **98 %** → more efficient!
- Column B: **91 %**



Average outlet TN over 140 days:

- Column A: **11.9 mg N/L** → removal **66 %**
- Column B: **4.3 mg N/L** → removal **87 %**

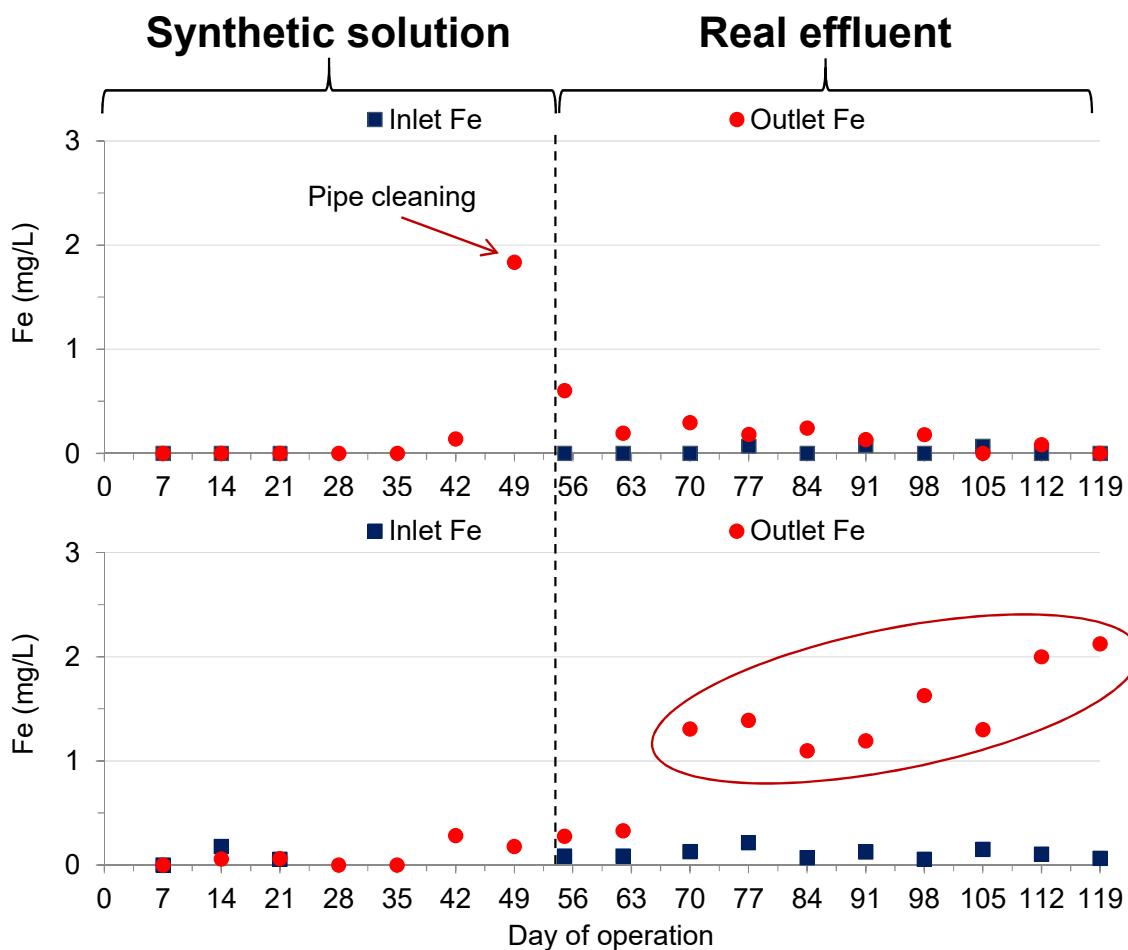
Column B shows higher TN removal:

- **Heterotrophic denitrification under anoxic conditions**

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Results & discussion: Fe release



Column A: aerobic

Synthetic solution (day 1-54):

- Outlet Fe: $0.28 \pm 0.69 \text{ mg Fe/L}$

Real effluent (day 55-120):

- Outlet Fe: $0.21 \pm 0.17 \text{ mg Fe/L}$

Column B: anoxic

Synthetic solution (day 1-54):

- Outlet Fe: $0.23 \pm 0.07 \text{ mg Fe/L}$

Real effluent (day 55-120):

- Outlet Fe: $1.26 \pm 0.61 \text{ mg Fe/L}$

➤ Fe release from CBR

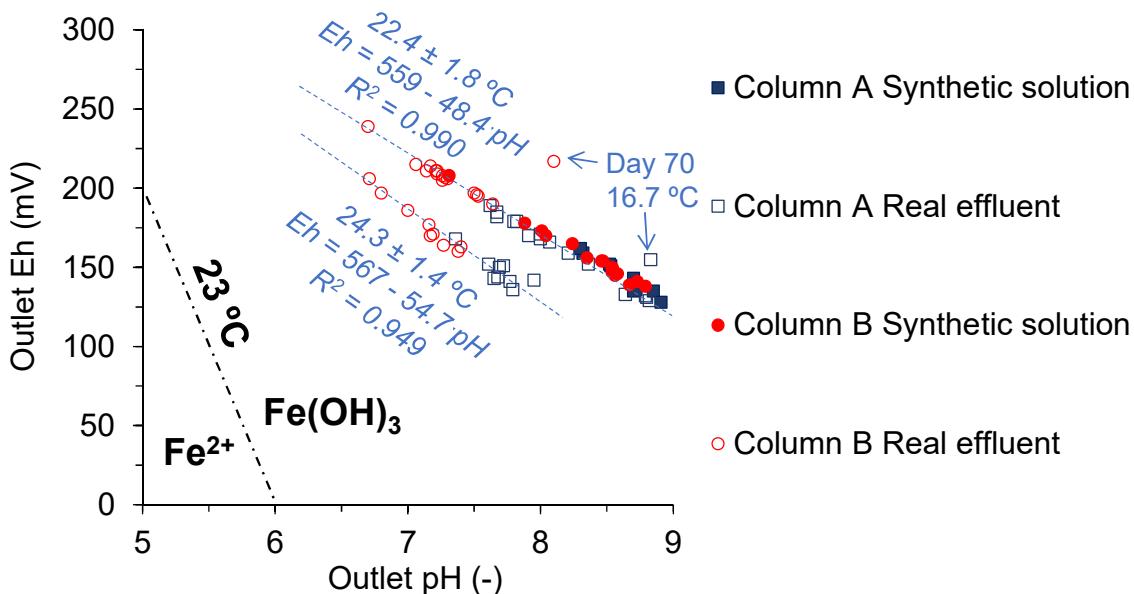
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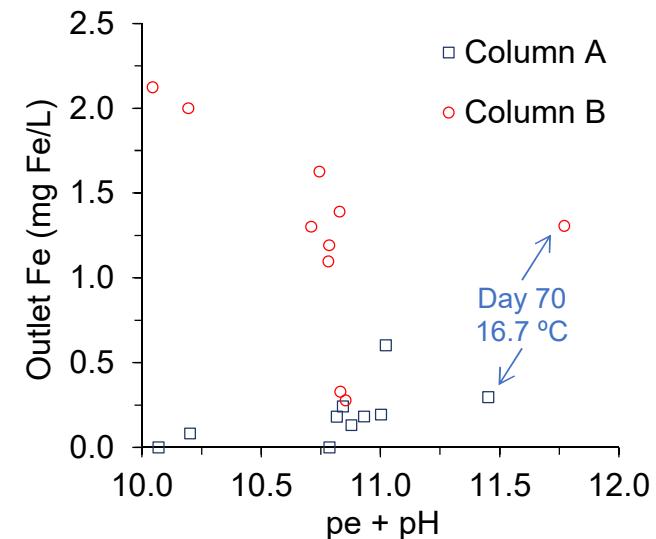
Results & discussion: Fe release

Simplified iron Poubaix diagram at 23°C:
outlet Eh and pH values of columns A and B.



- Linear correlation: Eh increases with decreasing pH;
- Eh and pH into the range of values that promote Fe(OH)₃;
- Eh and pH values alone cannot explain Fe mobilisation.

Outlet Fe concentrations of columns A and B as a function of pe + pH (days 55-140).



- pe + pH < 13.8: reducing media;
- Column B: outlet Fe increases with decreasing pe + pH below 11;
- Anoxic biofilm promoted Fe mobilisation.

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Material & methods: chemical extractions

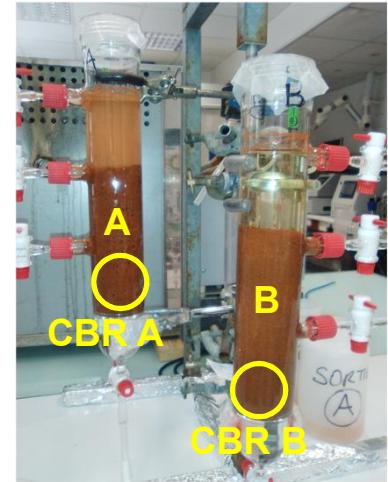
Chemical extractions: to identify main mechanisms of P removal.

Three different samples of CBR:

- Fresh CBR: CBR before the use to treat water;
- CBR A: CBR from the inlet of column A;
- CBR B: CBR from the inlet of column B.

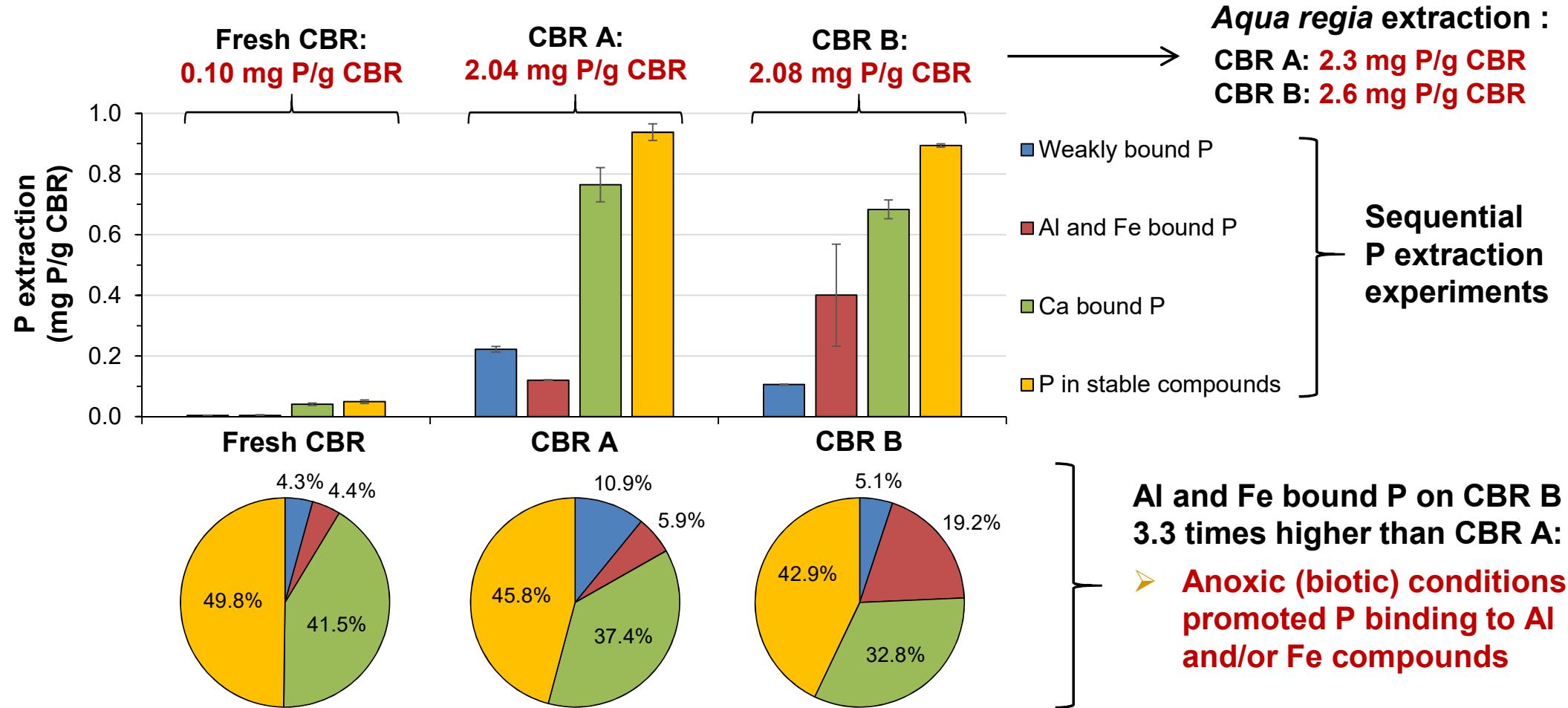
} After 140 days of column operation

1. **Aqua regia extractions (EN 13346, 2000): to determine total P content.**
2. **Sequential extractions (Moir et al., 1993; Barca et al., 2014): to quantify:**
 - i. Bicarbonate extractable P: weakly bound P;
 - ii. Hydroxide extractable P: leachable Al and Fe bound P;
 - iii. Diluted acid extractable P: leachable Ca bound P;
 - iv. Hot concentrated acid extractable P: P in stable residual compounds*.
3. **Amorphous Fe extractions (EN 12782-1, 2009): reactive Fe under amorphous form.**

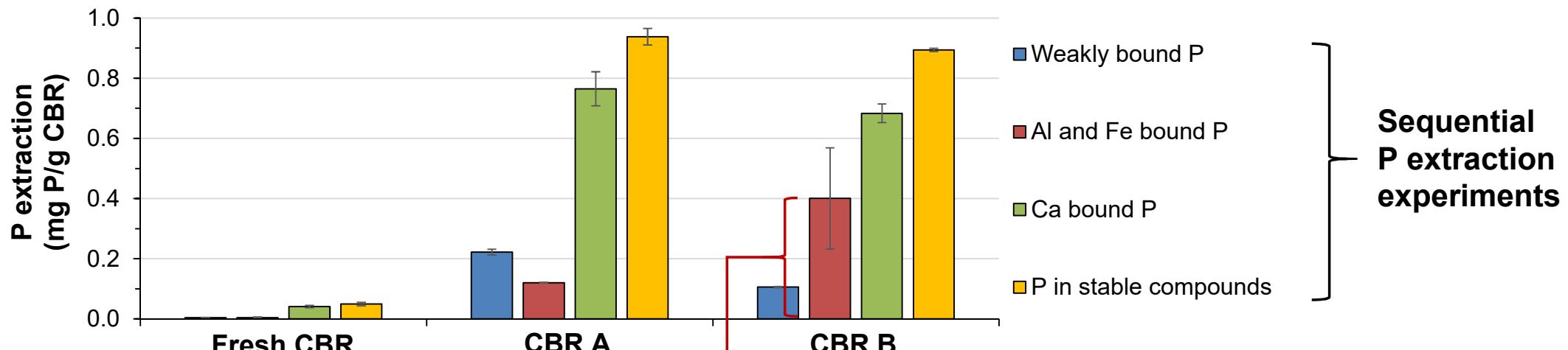


*Mainly attributed to: recalcitrant Ca-P crystals and/or organic P.

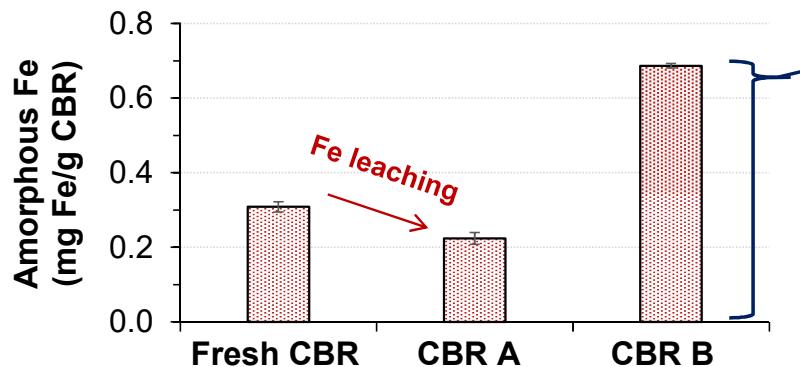
Results & discussion: P removal mechanisms



Results & discussion: P removal mechanisms



Amorphous Fe extraction experiments:



Amorphous Fe content of CBR B 3.1 times higher than CBR A:

- **Mobilization of stable Fe under anoxic (biotic) conditions:**
 - More Fe was available for P binding;
 - Fe-release from column B (anoxic).



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Conclusions

CBR is an efficient material to remove P from wastewater:

- High P retention capacity (> 4 g P/kg CBR);
- High P removal performance over the full period of 140 days of operation:
 - Column A (aerobic): 98.5%;
 - Column B (anoxic): 91.6.
- Almost neutral effluent pH (7-8);
- Good hydraulic conductivity.

Main mechanism of P removal:

- Precipitation of Ca-P complexes;
- P binding to Al and/or Fe compounds.

Anoxic (biotic) conditions can promote mobilization of Fe from CBR.

- The use of CBR filters appears to be particularly suitable as a tertiary treatment step to remove P from effluents with low organic load and under aerobic conditions.

*Thank you for your
attention!*

Barca et al., 2021. <https://doi.org/10.1016/j.jwpe.2020.101757>